

Taxonomic identification and diversity of effective soil microorganisms: towards a better understanding of this microbiome

Identificación taxonómica y diversidad de microorganismos efectivos del suelo: hacia un mejor entendimiento de este microbioma

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ABSTRACT

Soil microorganisms found in agricultural residues and the so-called efficient microorganisms (EM) are attractive for their potential applications and benefits in the bioremediation of complex ecosystems. However, the knowledge about *Who is doing what?*, as well as the trophic interaction in those communities that explain its benefits are limited; a better understanding of this microbiome is needed to explain its benefits. The objective of this research was to characterize the microorganisms isolated from two soil communities and the efficient microorganisms obtained in laboratory (EM16 consortium), taking into account physico-chemical characteristics, diversity, quantification, and taxonomic identification through microbiological and molecular techniques. A microbiological analysis was performed according to the morphological characteristics of the colonies as well as the study of the dynamics and taxonomic identification of the microbial populations through the TRFLP and Ion Torrent techniques. The diversity, dynamics, and taxonomic identification achieved in these studies showed the prospects for using these soil EM in bioremediation, considering the diverse metabolic pathways that these species have and their symbiotic interactive potential for biodegradation of lignocellulosic-resilient compounds. This study provides the first molecular characterization of the EM (EM16 consortium) and soil isolates from agricultural residues (sugarcane crop and bamboo field). The results suggest that the use of microbiological and molecular tools in a polyphasic approach allows the complete characterization of non-cultivable microorganisms that could contribute to sustainable environmental management and crop production.

Key words: molecular methods, soil microorganisms, agricultural residues, polyphasic approach.

RESUMEN

Los microorganismos del suelo que se encuentran en los residuos agrícolas y los llamados microorganismos eficientes (ME) son atractivos por su potencial aplicación y beneficios en la biorremediación de ecosistemas complejos. Sin embargo, el conocimiento sobre *¿Quién hace qué?*, así como la interacción trófica en esas comunidades que explican sus beneficios son limitados; se necesita una mejor comprensión de este microbioma que explique sus beneficios. El objetivo de esta investigación fue caracterizar los microorganismos aislados de dos comunidades de suelo y los ME obtenidos en el laboratorio (consorcio EM16), teniendo en cuenta las características físico-químicas, la diversidad, la cuantificación y la identificación taxonómica mediante técnicas microbiológicas y moleculares. Se realizó un análisis microbiológico según las características morfológicas de las colonias, así como el estudio de la dinámica e identificación taxonómica de las poblaciones microbianas mediante las técnicas TRFLP e Ion Torrent. La diversidad, dinámica e identificación taxonómica logradas en este estudio mostraron las perspectivas para uso de estos ME del suelo para la biorremediación, considerando las posibles rutas metabólicas que tienen estas especies y su potencial de interacción simbiótica para la biodegradación de compuestos lignocelulósicos resistentes. Este estudio proporciona la primera caracterización molecular de los ME (consorcio EM16) y de aislados del suelo procedentes de residuos agrícolas (cultivo de caña de azúcar y campo de bambú). Los resultados sugieren que el uso de herramientas microbiológicas y moleculares en un enfoque polifásico permite la caracterización completa de microorganismos no cultivables que podrían contribuir a la gestión ambiental sostenible y a la producción de cultivos.

Palabras clave: métodos moleculares, microorganismos del suelo, residuos agrícolas, enfoque polifásico.

Introduction

A wide diversity of microbial species interacts in natural ecosystems, contributing to waste degradation (Sun *et*

al., 2013; Azman *et al.*, 2015; Raja *et al.*, 2017). The use of agricultural crop residues and agro-industrial waste would be attractive alternatives as a source of renewable energy

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because they close production cycles in a circular economy approach. However, despite the diversity of microorganisms involved in the decomposition of organic matter, the efficient biodegradation of these residues remains a challenge for bioethanol and biogas production due to the high percentages of lignocellulosic material, whose polymers are difficult to decompose (Cortés, 2016; Widjaja *et al.*, 2016; Amin *et al.*, 2017). The search continues, therefore, for microorganisms that naturally contain this lignocellulolytic capacity that can be exploited biotechnologically.

Teuro Higa developed in 1986 the technology of so-called efficient microorganisms (EM) (Allahverdiyev, Kırdar *et al.*, 2011; Higa & Parr, 2013) that are isolated from undisturbed soils and ecosystems. EM technology has been applied in agriculture and forestry in more than 80 countries worldwide (Calero Hurtado, Pérez Díaz *et al.*, 2019; Tanya & Leiva-Mora, 2019) with beneficial impacts. Microbial community compositions related to anaerobic EM consortia have not been characterized, but previous studies have identified phylogenetic groups, including Actinomycetales and fermenting fungi, photosynthetic bacteria, yeasts, and lactic acid bacteria (Namasivayam *et al.*, 2014; Joshi *et al.*, 2019).

Soil microorganisms, including EM consortia, play a crucial role in agricultural production and environmental decontamination, as they allow the availability of assimilable nutrients for plants during the biodegradation of organic matter, where phytohormones and secondary metabolites are released (López-Dávila, Gil Unday *et al.*, 2017; Calero Hurtado, Quintero Rodríguez *et al.*, 2019; Joshi *et al.*, 2019; Tanya & Leiva-Mora, 2019; Castro *et al.*, 2022). However, physico-chemical characteristics and microbiological composition of EM consortia should be clarified in order to understand their beneficial effects on agricultural production (Calero Hurtado *et al.*, 2020). The beneficial uses of EM consortia include restoration of damaged soils after application of chemical fertilizers and pesticides (Alvarez *et al.*, 2018), manufacturing of fermented fertilizers (biofertilizers) (Calero Hurtado, Quintero Rodríguez *et al.*, 2019), processing of organic waste, and wastewater treatment (Allahverdiyev, Atilla *et al.*, 2011). It has been widely reported that *Lactobacillus* species are prevalent in the EM consortia that is the basis for their use as a biofertilizer to enhance agricultural production (Blainski *et al.*, 2018; Daranas *et al.*, 2018; Quattrini *et al.*, 2018; Naik *et al.*, 2019; Abd El-Mageed *et al.*, 2020; Muhialdin *et al.*, 2020).

Therefore, an important strategy would be to determine not only the main physico-chemical characteristics but

also which of them contribute to the improvement of soil properties, growth of plants, and other important environmental processes (Higa & Parr, 2013). The EM consortia works synergistically to release beneficial substances such as vitamins, hormones, enzymes, organic acids, bioactive minerals, and various antioxidants when they come into contact with the soil organic matter (Allahverdiyev, Kırdar *et al.*, 2011). In addition, the EM consortia helped to improve the soil pH and increase the action of mineral nutrients, hormones, and other metabolites that accelerate decomposition of organic wastes and increase formation of biogas (López-Dávila, Gil Unday *et al.*, 2017). Therefore, the hydrolytic and fermentative activities of the EM consortia could be potentially used as a pre-treatment or biostimulant in the agricultural waste bioconversion processes into bioenergy, for instance in bioethanol or biogas production.

The microbial dynamics of natural ecosystems (soils) as well as artificial ecosystems (anaerobic reactors), can be very complex due to the wide diversity of species and metabolic interactions required to degrade the organic matter. Microbiological characterization of these ecosystems using culture-dependent techniques are insufficient since only 0.1-10% of the bacteria in the environment can be cultivated (Nobu *et al.*, 2015; Saw *et al.*, 2015; Jiménez-Hernández *et al.*, 2021). Hence, it is necessary to apply molecular methods that allow the analysis of taxonomic diversity and spatial structure of complex microbial communities to identify specific microbial populations in their natural habitat as well as to predict existing metabolic interactions.

Previous studies for microbiome monitoring employed molecular techniques based on 16S rRNA gene analysis such as denaturing gradient gel electrophoresis (DGGE), terminal restriction fragment length polymorphisms (TRFLP), single-strand conformation polymorphism (SSCP) or capillary sequencing of selected genes by the Sanger method (Zoetendal *et al.*, 2008). In complex microbial ecosystems, these methods provide an incomplete composition of the microbial community, showing only the existence of certain groups. The development of metagenomic approaches based on next-generation sequencing (NGS) has been one of the great advances in molecular biology (Zhou *et al.*, 2015; Sanz & Köchling, 2019). The combination of different technologies seems to be imperative to reach the goals. Techniques complement each other, being the weaknesses of some and the strengths of others, especially when the study requires the identification of species to elucidate predominant metabolic pathways (Carabeo-Pérez *et al.*, 2019).

Since 2014, in the Microbiology Laboratory of the University of Sancti Spíritus (Universidad de Sancti Spíritus “José Martí Pérez” - UNISS, Cuba), EM consortia have been developed, taking into account the methodology proposed by Teuro Higa and modified by Olivera-Viciedo *et al.* (2014). These studies were focused on the application of EM as a biostimulant to soils and plants. However, a detailed characterization of this microbial consortium has not been performed.

Agricultural soils in Cuba are currently in unfavorable conditions due to intensive cultivation and inadequate management (Febles-González *et al.*, 2014). Most of the agricultural soils in this overexploited condition are located in the central region of the country that is a major producer of sugarcane (ONEI, 2020). These are brown carbonate soils with a pH between 5.0 and 8.5. In addition, they have high N, P and K content due to fertirrigation with wastewater from sugar factories and distilleries (Crespo *et al.*, 2018). However, the diversity and dynamics of the microbial populations present in this soil and how they contribute to the degradation of agricultural residues were not found in previous studies.

There are also extensive areas of soil in Cuba that have very little anthropogenic action and are of special interest. These are generally forested areas covered by bamboo species (*Phyllostachys reticulata* [Ruprecht] Koch) and marabu (*Dichrostachys cinerea* [L.] Wight & Arnott). In the case of bamboo, they are generally found in forested areas near rivers or canals with alluvial soils (non-carbonate) (Oca-Risco *et al.*, 2014), and marabu species can germinate in any type of soil. No further molecular characterization of these virgin ecosystems was done in this study.

The objective of this research was to characterize the microorganisms isolated from two soil communities (sugar cane crop and bamboo forest) and the efficient microorganisms obtained in the laboratory (EM16 consortium), taking into account physico-chemical characteristics, diversity, quantification, and taxonomic identification through molecular techniques.

Materials and methods

Obtaining isolates of EM consortium and agricultural soils

The EM consortium was produced in the Microbiology Laboratory of the University of Sancti Spíritus (UNISS, Cuba) pilot plant (20 L fermenter), following a methodology similar to that proposed by Olivera-Viciedo *et al.* (2014).

Production was carried out as follows: 5 L of whey was mixed with 5 L of “C molasses” from a sugarcane factory. This mixture was added in layers to 10 kg of decomposing foliage and soil obtained from virgin forests undisturbed by human activity in the vicinities of the Zaza River of the Southern region of Sancti Spiritus province. Nine kg of corn flour was added to the mixture which was homogenized, compacted, and sealed in an anaerobic tank. The mixture (solid EM material) was kept at 28°C in the dark for 21-25 d for fermentation. After this time, 2 kg of this solid EM material was diluted in 18 L of distilled water containing 1 L of whey and 1 L of “C molasses” to obtain a total volume of 20 L of liquid bioproduct. The culture was kept hermetically sealed for 7 d under the same conditions described above, to avoid disturbing the fermentation process. The consortium obtained was called EM16.

To obtain the culture of soil microorganisms, 10 samples of 20 g were collected from two different agroecosystems: i) 10 cm of soil beneath the leaf floor of sugarcane field 7 d after harvest and ii) 10 cm of soil beneath the leaf floor of 5-year-old bamboo (*Bambusa vulgaris* Schrader) field (brown soils with carbonates and alluvial). Each soil sample was individually suspended in 350 ml of peptone cellulose solution (PCS medium: 0.1% yeast extract, 0.5% peptone, 0.2% CaCO₃, 0.5% NaCl, 0.5% cellulose, pH 7.0) for 7 d at 30°C in static culture flasks to isolate microbial strains capable of degrading lignocellulosic substrates.

Physico-chemical characterization of EM

The EM16 consortium and the soil isolates grown for 7 d in PCS were characterized by the following physico-chemical parameters: dry matter (DM), volatile solids (VS), ash, and pH, according to the 23rd Standard Methods for the Examination of Water and Wastewater (Baird *et al.*, 2017).

Morphological analysis (growth in plates)

To identify the cultural bacteria, NA (nutrient agar) medium was used, employing surface spreading plate technique, with 0.1 ml of original samples and dilutions 10⁻¹, 10⁻² and 10⁻³ as inoculum. Petri plates were incubated at 37°C for 96 h, after which it was possible to differentiate the colonies by their morphological characteristics in the culture.

DNA extraction

Genomic DNA was extracted from each sample (triplicate aliquots of 200 mg) using the FastDNA[®] SPIN Kit for Soil and the FastPrep[®] Instrument (MP Biomedicals, USA) according to the manufacturer’s guidelines. The quantity and purity of DNA were determined photometrically using a NanoDrop 2000/2000c (Thermo Scientific, USA)

according to the manufacturer's guidelines. Isolated DNA was stored at -20°C until further processing.

Amplification of bacterial 16S rRNA gene sequences

Genes encoding for bacterial 16S rRNA (*rrs*) were amplified using the PCR primers 27f (5'-AGAGTTTGATCMTGGCT-CAG-3') (Lane, 1991; Sipos *et al.*, 2007) and 1492r (5'-TACGGYTACCTTGTTACGACTT-3') (Weisburg *et al.*, 1991; Després *et al.*, 2007). The PCR reaction mix contained 1X *Taq* buffer, 2 mM MgCl₂, 0.2 mM of each deoxynucleoside triphosphate, 0.4 μM of each primer, 1 μl (10 ng approximately) of template DNA and 1 U of *Taq* DNA polymerase for a final volume of 25 μl. The amplification was performed with an initial denaturation step for 1 min at 94°C, followed by 34 cycles of denaturation at 94°C for 1 min, annealing at 57°C for 1 min, extension at 72°C for 2 min and a final extension stage at 72°C for 3 min.

Chemicals and enzymes were provided by Fermentas (St. Leon-Rot, Germany), Promega (USA) Corporation and Molecular Biology (Thermo Scientific, USA). The correct length of the amplicons was verified by 1% agarose gel electrophoresis, using the GeneRuler® 1kb Plus DNA Ladder (Thermo Scientific, USA). To minimize the risk of any PCR bias, three parallel PCR reactions were performed for each of the three parallel DNA extracts obtained from each environmental sample. The concentration of the amplified 16S rRNA gene fragments was measured using the Qubit4® fluorometer (Invitrogen, Thermo Scientific, USA).

Terminal Restriction Fragment Length Polymorphism analysis (TRFLP)

For the bacteria diversity study, TRFLP analysis was carried out following the protocol proposed before by Rademacher *et al.* (2012). The primers 27f (5'-AGAGTTTGATCMTGGCTCAG-3') (Lane, 1991; Sipos *et al.*, 2007), labeled at the 5' terminal end with Indodicarbocyanin (Cy5) and 926r (5'-CCGTCAATTCMTTTRAGTTT-3') (Weisburg *et al.*, 1991; Després *et al.*, 2007) were used. The three independent amplification products, based on the same DNA template, were pooled and purified by applying the PureLink® PCR Purification Kit (Invitrogen, Thermo Scientific, USA). The concentration of purified products was measured using the Qubit4® fluorometer (Invitrogen, Thermo Scientific, USA).

After purification, the PCR products were digested with *Msp*I and *Hin*6I following the manufacturer guidelines for each restriction enzyme. The digestion fragments were electrophoretically separated and detected by fluorescence using a GenomeLab® GeXP Genetic Analysis System

(Beckman Coulter, Krefeld, Germany). The data obtained were analyzed using the DataConnect® software (Applied Biosystems interface-Thermo Fisher) considering the size calculation of the detected terminal restriction fragments (TRFs) based on the migration time of the applied size standard. All fragments with a sequence length between 60 and 640 bp were used for further analyses with the T-Rex software package, available online (<http://trex.biohpc.org>). The identification of "true" peaks by distinguishing baseline "noise" from signals of fluorescently labeled fragments, as well as the alignment of TRFs with a threshold of 0.5, was based on the evaluation of the peak height. In the last evaluation step, TRFs were visualized by their relative distribution, considering that TRFs with a relative abundance lower than 2% were removed from the analyses.

Bacterial 16S rRNA gene sequencing by Ion Torrent

To amplify the 16S hypervariable regions, the Ion 16S® Metagenomics Kit (Life Technologies, Thermo Scientific, USA) was used, on the Ion Torrent Personal Genome Machine® (PGM, Ion Torrent) platform, following the manufacturer guidelines. The remaining PCR reaction was purified using the PureLink PCR Purification Kit (Life Technologies, Thermo Scientific), and the concentration of the purified amplicons was assessed with the Qubit4 fluorometer and Qubit® dsDNA High Sensitivity Assay Kit (Life Technologies, Thermo Scientific). The DNA library construction, quantification, template preparation, and sequencing were developed according to the standard protocol, well-described by Adamiak *et al.* (2018). Metagenomic data was evaluated using the Ion Reporter® software (Invitrogen, Thermo Scientific, USA), showing the taxonomic distribution of the target microbial community, based on the software database. The raw data of sequences have been deposited in the Sequence Read Archive (SRA) of the National Center for Biotechnology Information (NCBI) under the Bioproject accession number PRJNA764521 (<http://www.ncbi.nlm.nih.gov/bioproject/764521>).

Analysis of the diversity and organization of the microbial communities

The bacteria richness (Rr) considering more than 1% of abundance was calculated by the methodology proposed by Marzorati *et al.* (2008), based on the number of OTUs obtained in each electropherogram or cluster. The Shannon-Weaver diversity index (H) (Shannon & Weaver, 1963) was used to evaluate and compare this diversity, and the evenness index (Pielou index, H/H_{max}) of the microbial community in each sample was calculated as the quotient of the Shannon index and the potential number of species

in the sample (Hmax). This index describes the uniformity of the distribution of individuals in the community (Pielou, 1966).

Results and discussion

Physico-chemical characterization of EM16

The EM16 consortium presented low contents of dry matter (DM), which ranged from 0.90% to 1.57% for fresh matter (FM). Also, more than 50% of DM are volatile solids (VS) (0.51-0.95 of VS (% FM)), and 56.8-60.67 VS (% DM) is attributable to the biomass content. During the 7 d anaerobic fermentation process, biomass production is limited due to the anaerobic condition (Madigan *et al.*, 2019). In addition, the dilution during the EM preparation described above (EM and agricultural soils isolates obtention) could contribute to this measure. Dai *et al.* (2016) also described a low organic matter content in EM bioproducts. The EM16 consortium had acid properties (pH = 4.31-4.46), agreeing with other reports where the pH ranged from 3.1 to 4.7 (López-Dávila, Calero Hurtado *et al.*, 2017; Calero Hurtado *et al.*, 2020). In the EM16 consortium, the low pH might be due to: i) the use of substrates with low pH like whey and molasses (Núñez-Caraballo *et al.*, 2019; González-Herrera *et al.*, 2021) and ii) the volatile organic acids formed during the fermentation processes (Xiong *et al.*, 2012; Dai *et al.*, 2016). In addition, the low pH reported by Calero *et al.* (2020) could be linked to the assimilation of some important mineral nutrients, such as N, P, K, and Ca that help to improve soil properties and plant growth. Also, knowledge of this characterization could help to identify substances important for use in agricultural production and other environmental practices.

Morphological analysis

The EM16 sample did not showed colony formation on plate growth (in aerobic condition). This result may be related to the lower relative abundance of the bacterial species in this sample with respect to the other isolated soil or that these microorganisms were not capable of growing under these conditions. EM consortium is frequently composed mainly of *Lactobacillus* species (Tanya & Leiva-Mora, 2019) that are facultative anaerobic bacteria difficult to grow in a Petri dish. On this basis, Calero *et al.* (2020) reported that the EM consortium is composed of some species of *Lactobacillus bulgaricum*, *Bacillus subtilis* and *Saccharomyces cerevisiae* that represent microbial diversity. Therefore, in the EM16 consortium, some of these species should also be identified.

In contrast, in the undiluted soil isolate samples, bacterial growth exceeded 300 colonies after 24 h of incubation;

those plates are reported as uncountable. However, the diluted samples (10^{-2} and 10^{-3}) showed colony growth. After the incubation period, it was possible to differentiate the bacteria colonies according to the different morphological characteristics in the culture.

In sugarcane soil isolates, bacterial colonies were observed mostly in a circular, irregular, or rhizoid shape, with rounded, wavy, or lobed edges and different surfaces such as flat, convex, accumulated, and umbilical. Bamboo soil isolated colonies with irregular or circular shapes with flat surfaces and wavy edges predominated.

The observed morphological diversity demonstrated that a great variety of metabolic pathways may also exist in soil microorganisms. This physiological diversity allows the microbial community to maintain the balance of agroecosystems. To guarantee soil health, decomposition of organic matter, crop yields, availability of the nutrients, etc., the microorganisms establish a range of ecological interactions that make the microbial populations develop or not (Vilatuña, 2019). These relationships are poorly studied and should be considered in research on soil microbiota (Goncharov & Tiunov, 2014; Jacoby *et al.*, 2017; Erktan *et al.*, 2020). When the microbial culture is performed, the growth of the entire community will not be observed because, perhaps, the growth of some species depends on others and once isolated in a different culture medium it does not grow; perhaps the new culture medium does not provide the necessary nutrients or perhaps the dilution factor used affected growth.

Traditional phenotypic identification schemes based on “observable” colony characteristics, such as morphology, development, and biochemical and metabolic properties, are essential as diagnostic methods for microbial community studies. However, these are insufficient since, from these methods, it is only possible to observe the growth of those facultative anaerobic microorganisms; but when it comes to environmental samples or samples taken from anaerobic systems, the use of other molecular techniques that contribute to quantification and microbial identification is recommended. It is important to determine exactly the relationship of the form and physiology of those microorganisms because it would help to understand their importance and function within the processes and how to manage them better.

Bacterial community structure and dynamics

In EM16 samples, 49 terminal fragments belonging to the Bacteria Domain were detected by TRFLP analysis.

Terminal fragments (TFs) with 62, 181, 404, 476, 546, 571, 572, 575, 579, 582, 585, 597, 600, and 610 bp showed high relative abundance of more than 2% (Fig. 1). From the sugarcane straw soil isolated samples, 94 TFs were retrieved, where the fragments of 192, 194, 204, 225, 276, 282, 451, 453, 472, 505, 508, 525, 546, 571, 575, 585, and 597 bp showed more than 2% relative abundance. From the bamboo soil isolated samples, 83 TFs were obtained, where the TFs of 152, 192, 194, 204, 225, 276, 281, 282, 451, 453, 472, 505, 508, 525, 546, and 579 bp showed a relative abundance more than 2% (Fig. 1). When the total communities are explored, including the TFs with a relative abundance below 2%, 13 TFs were found in both isolated soil samples (*i.e.*, 192, 194, 204, 225, 276, 282, 451, 453, 472, 505, 508, 525, and 546 bp). That is, probably, because they are endemic microbiota from these niches (brown carbonate soils).

According to Ion Torrent identification results from the Bacteria Domain (Tab. 1), five phyla were detected in the samples studied.

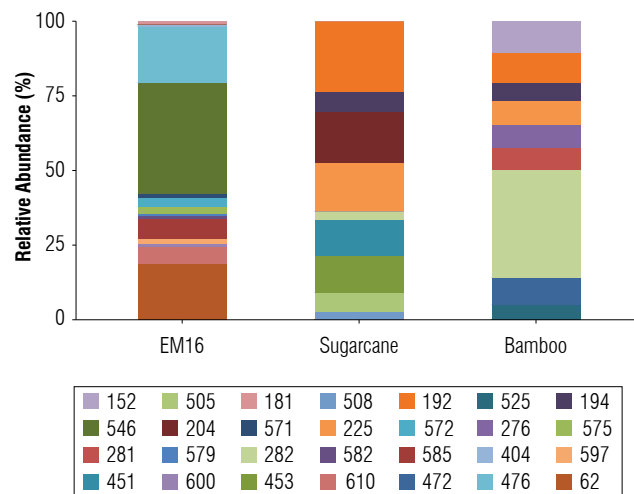


FIGURE 1. Relative abundance of bacteria terminal restriction fragments and their sizes (bp) in samples of EM16 and soil isolates (sugarcane crop and bamboo field). The relative abundance of each fragment was determined based on the height of its maximum of fluorescence in relation to the total of the heights of all the peaks detected in the sample.

TABLE 1. Taxonomic distribution of the bacteria domain in samples of EM16 consortium and soil isolates identified by Ion Torrent technique.

Phylum	Class	Order	Family	% of mapped reads in EM16	% of mapped reads in sugarcane	% of mapped reads in bamboo		
Actinobacteria	Actinobacteria	Actinomycetales	Microbacteriaceae	0.09	0	0		
			Propionibacteriaceae	0.16	0	0		
Bacteroidetes	Bacteroidia	Bacteroidales	Bacteroidaceae	0	0	0.36		
			Porphyromonadaceae	0	0	15.53		
	Flavobacteriia	Flavobacteriales	Flavobacteriaceae	0	1.86	0.42		
	Sphingobacteriia	Sphingobacteriales	Chitinophagaceae	0.11	0	0		
			Sphingobacteriaceae	0	0.2	1.68		
Firmicutes	Bacilli	Bacillales	Alicyclobacillaceae	0	0.38	0		
			Bacillaceae	0	1.59	10.62		
			Sporolactobacillaceae	92.25	0	0		
			Paenibacillaceae	0	2.98	2.75		
			Lactobacillales	Lactobacillaceae	3.25	0	0	
			Christensenellaceae	0	0	0.21		
			Clostridiaceae	1.62	22.13	16.22		
Firmicutes	Clostridia	Clostridiales	Clostridiales Family XI. Incertae Sedis	0	4.36	3.17		
			Eubacteriaceae	0	0.63	0		
			Lachnospiraceae	0	1.68	0.78		
			Oscillospiraceae	0	0	0.1		
			Peptococcaceae	0	2.36	2.69		
			Peptostreptococcaceae	0	12.76	9.05		
			Ruminococcaceae	0.08	17.79	8.09		
			unclassified Clostridiales	0	0	0.91		
			Erysipelotrichia	Erysipelotrichales	Erysipelotrichaceae	0	0	0.33

to be continued

Phylum	Class	Order	Family	% of mapped reads in EM16	% of mapped reads in sugarcane	% of mapped reads in bamboo	
Proteobacteria	Planctomycetia	Candidatus Brocadiales	Candidatus Brocadiaceae	0.05	0	0	
			Caulobacterales	Caulobacteraceae	0.08	9.41	0.47
	Alphaproteobacteria	Rhizobiales		Bradyrhizobiaceae	0.1	0	0
				Acetobacteraceae	0.32	0.45	0.92
		Rhodospirillales		Rhodospirillaceae	0	5.44	5.87
				Alcaligenaceae	0	0	0
	Betaproteobacteria	Burkholderiales		Burkholderiaceae	1.46	0	0
				Comamonadaceae	0.16	0.8	2.34
	Epsilonproteobacteria	Campylobacterales		Campylobacteraceae	0.06	0	0
	Gammaproteobacteria	Aeromonadales		Aeromonadaceae	0	0	0
				Enterobacteriaceae	0.1	7.77	6.39
		Pseudomonadales		Moraxellaceae	0	0.93	1.5
				Pseudomonadaceae	0.04	4.39	8.3
		Xanthomonadales		Xanthomonadaceae	0.08	0.22	0
Total				100.00	98.13	98.7	

Almost 97% of the microbial community in EM16 was represented by the Firmicutes phylum. Other species were detected belonging to the phyla Proteobacteria (2.4%), Actinobacteria (0.25%), Bacteroidetes (0.11%), Planctomycetes (0.05%) (Tab. 1 and Fig. 2). A similar composition of microbial structure was identified in a thermoacidophilic EM consortium with six predominant phyla (Proteobacteria, Firmicutes, Chloroflexi, Bacteroidetes, Actinobacteria, and Acidobacteria) (Henry *et al.*, 2020). The result of this study showed that the microbial community in the EM16 consortium is favored by the formation of *Sporolactobacillus* species (family Sporolactobacillaceae 92.25%) (Fig. 2).

The members of *Sporolactobacillus* genus are catalase-negative, microaerophilic, spore-forming, homofermentative, lactic acid-producing species and require mainly carbohydrate for growth (Chang *et al.*, 2008). Although *Sporolactobacillus* species are typically isolated from soil, they are occasionally isolated from fermented or spoiled foods (Yanagida *et al.*, 1987; Fujita *et al.*, 2010). The presence of these species was expected in the EM16 consortium due to the composition of the substrate (made in part by whey) considering that *Sporolactobacillus*, as well as *Lactobacillus*, are common in all types of dairy products (Madigan *et al.*, 2019). The presence of lactic acid bacteria related to the breakdown of cellulolytic and lignified organic materials has also been reported in EM (Salminen & von Wright, 2004; Joshi *et al.*, 2019).

The families Lactobacillaceae (3.14%) and Clostridiaceae (1.62%) were less abundant in this consortium (Fig. 2). The presence of *Clostridium* species is essential for cellulose degradation (Madigan *et al.*, 2019), therefore they could be used to improve the degradation of lignocellulosic residues. On the other hand, *Lactobacillus* spp. have potential effects for agricultural needs, so they can enhance soil properties since they show an antagonistic effect against different phytopathogenic agents in the soil, mainly due to the decrease in pH, production of peptides with antimicrobial activity such as class I bacteriocins, and nisin that is very active against gram-positive bacteria (Tanya & Leiva-Mora, 2019). In addition, *Lactobacillus* spp. can enhance the trophic interactions (Quattrini *et al.*, 2018; Naik *et al.*, 2019), biotic and abiotic responses (Tsuda *et al.*, 2016; Blainski *et al.*, 2018; Abd El-Mageed *et al.*, 2020; Muhialdin *et al.*, 2020), and plant growth and productivity (Daranas *et al.*, 2018; Quattrini *et al.*, 2018). The presence of these species is, therefore, essential in the EM16 consortium to enrich the soil microbiome and its functionality during the cycle of assimilation and de-assimilation of nutrients.

The taxonomic groups identified in EM16 are similar to other EM consortia (Alvarez *et al.*, 2018; Naik *et al.*, 2019); however, the diversity and richness of species depend on the environmental conditions and the substrate from which they come. The EM16 consortium could be an efficient alternative to use in sustainable agriculture since it can promote microbial consortia interaction in the soil and

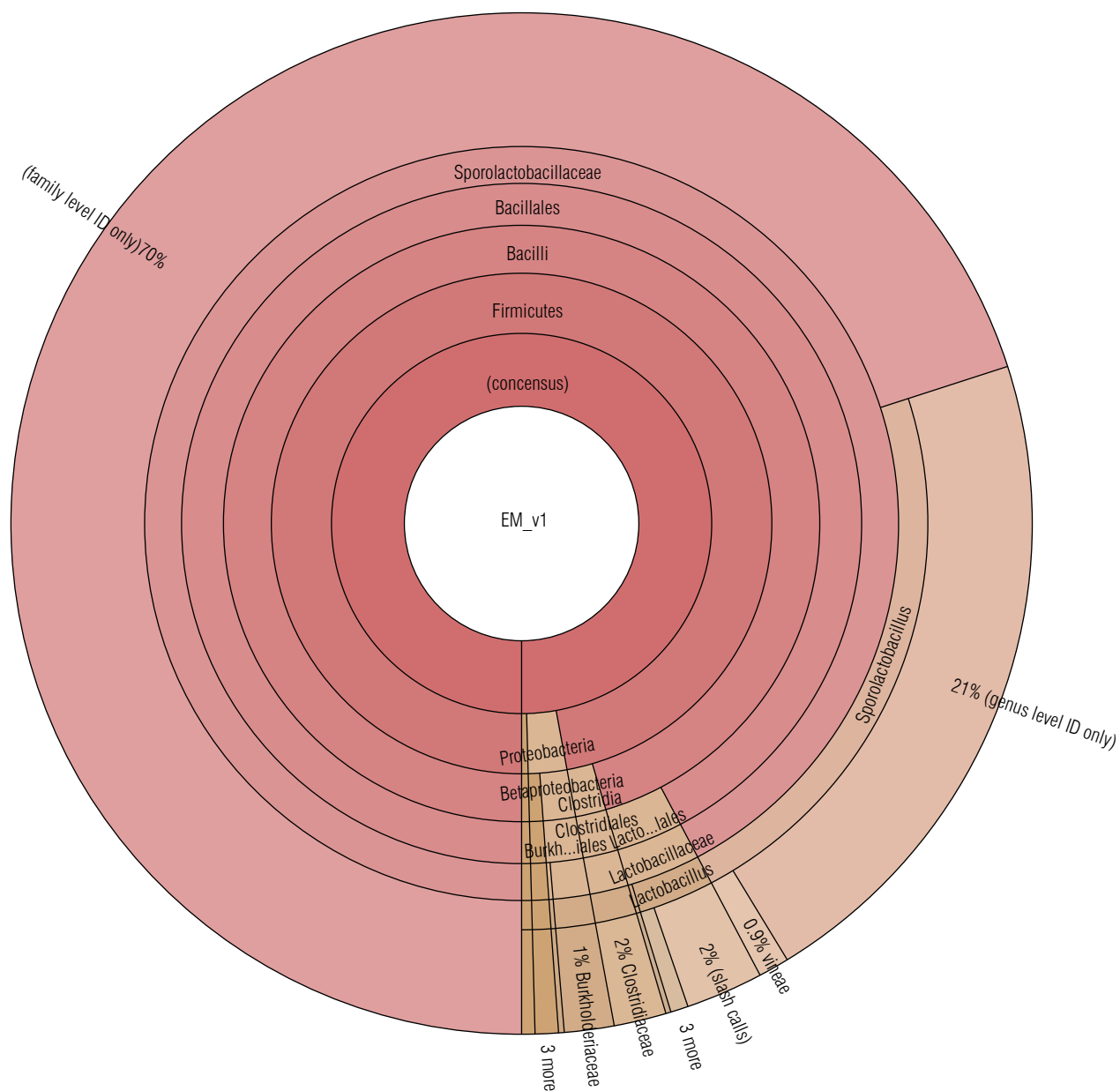


FIGURE 2. Krona plot of identified bacteria in EM16 sample, through the Ion Torrent technique, by using the Ion Reporter® software. The Krona plot shows the taxonomic distribution from phylum to species level and the associated abundance based on the percentage of mapped reads taking into account the NGS repositories based on NCBI database.

degrades organic components at the same time it supports a better micro or macronutrient assimilation by roots.

The sugarcane soil sample was represented by the Firmicutes (67.06%), Proteobacteria (29.64%), and Bacteroidetes (3.38%) phyla (Tab. 1). The most abundant families were Clostridiaceae (22.13%), Ruminococcaceae (17.59%), and Peptostreptococcaceae (12.76%) (Fig. 3).

A similar distribution of phyla was detected in bamboo soil samples, represented by Firmicutes (55.52%), Proteobacteria (26.28%), and Bacteroidetes (18.17%) (Tab. 1). The most abundant families were Clostridiaceae (16.22%), Porphyromonadaceae (15.53%), and Bacillaceae (10.62%) (Fig. 4).

In the sugarcane and bamboo soil samples (Figs. 3-4), where bacterial species belonging to the phylum Firmicutes



FIGURE 3. Krona plot of identified bacteria in sugarcane isolated soil sample, through the Ion Torrent technique, by using the Ion Reporter® software. The Krona plot shows the taxonomic distribution from phylum to species level and the associated abundance based on the percentage of mapped reads taking into account the NGS repositories based on the NCBI database.

predominated, more than 10 different species of the *Clostridium* genus were identified to the species level. These are fermentative bacteria, strictly anaerobic with great metabolic diversity, gram-positive, endospore-forming, generally fermenting sugars or amino acids usually found in soil samples (Madigan *et al.*, 2019). Besides, in the Clostridiales order, with an abundance greater than 1.5%

the genera *Ruminococcus* and *Sedimentibacter* have been identified in both soil samples.

Ruminococcus species are defined as strictly anaerobic, gram-positive, non-motile cocci that do not produce endospores and require fermentable carbohydrates for growth (Goodfellow *et al.*, 2009). They have been detected

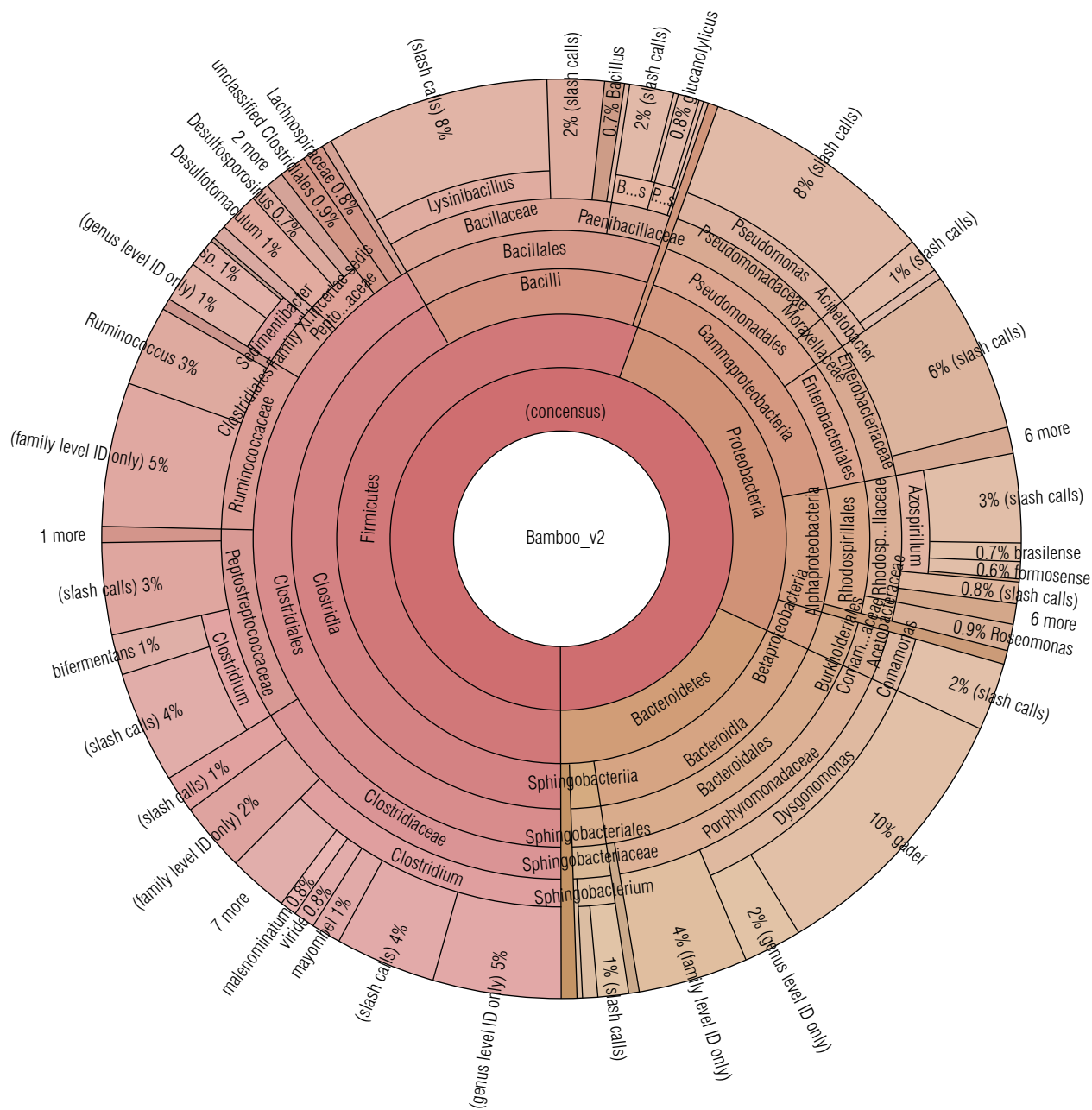


FIGURE 4. Krona plot of identified bacteria in bamboo isolated soil samples using the Ion Torrent technique, using the Ion Reporter® software. The Krona plot shows the taxonomic distribution from the phylum to species level and the associated abundance based on the percentage of mapped reads taking into account the NGS repositories based on the NCBI database.

in greatest abundance in samples taken from herbivorous and omnivorous animal hosts, but a few have been found in environmental sources, generally in host-associated samples (La Reau *et al.*, 2016). Some species are cellulolytic, playing an important role in the degradation of particulate substrates (Flint *et al.*, 2008). Other members of family Ruminococcaceae are numerically abundant in the human intestinal tract, like bacterial species capable of degrading crystalline cellulose (Chassard *et al.*, 2012; Morais *et al.*,

2016). Others are non-cellulolytic and use polysaccharides like resistant starches (Ze *et al.*, 2012) or selectively use various plant hemicelluloses (Wegmann *et al.*, 2014).

Species of the genus *Sedimentibacter* have been commonly isolated from anaerobic microbial communities (Breitenstein *et al.*, 2002; Woo *et al.*, 2004; Imachi *et al.*, 2016). Cells are slightly curved rods, motile, gram-positive and spores might be formed. Growth is strictly anaerobic and

requires yeast extract and is supported by the fermentation of pyruvate or amino acids in a Stickland-type reaction. Hydrogen is not produced and carbohydrates are not fermented (Breitenstein *et al.*, 2002).

Members of the genus *Pseudomonas* (class Gammaproteobacteria) were also identified in soil samples from sugarcane and bamboo. These species show great metabolic diversity and consequently are widely distributed in nature. Some species are pathogenic for humans, animals, or plants. The metabolism of *Pseudomonas* is typically respiratory with oxygen as the terminal electron acceptor; but some species can also use nitrate as an alternative electron acceptor and carry out oxygen-repressible denitrification (dissimilatory reduction of nitrate to N₂O or N₂), allowing growth to occur anaerobically (Palleroni, 2015). They are typically oxidase and catalase-positive, do not form spores, with no gas formation from glucose, and perform catabolism of the carbohydrates by the alternative Entner-Doudoroff pathway and the cycle of tricarboxylic acids (Madigan *et al.*, 2019).

The microbial diversity found in bamboo and sugarcane soil isolates could be used to enrich the EM16 consortium to achieve a more complete bioproduct (with a broader metabolic capacity).

The phylogenetic analysis coincides with the results of the TRFLP, where some fragments were detected in both soil samples. In general, species richness, diversity and the evenness index allows a comparison of isolated soil and EM16 consortium samples as shown in Table 2, considering the Ion Torrent and TRFLP analyses results. As expected, the richness, diversity, and Pielou index were higher in isolated soil samples compared to EM16 samples (Tab. 2) due to the dynamic conditions and the diversity of nutrients in these natural ecosystems. Despite the low diversity in the EM16 samples, it is worth noting the uniformity of the community with a coefficient of over 0.5.

TABLE 2. Richness, diversity, and evenness indexes calculated for each sample considering the Ion Torrent and TRFLP (in parenthesis) analysis results. Species with a relative abundance more than 1% were considered.

Index	EM16	Sugarcane	Bamboo
Rr	4 (8)	25 (9)	22 (9)
H	0.73 (1.64)	2.89 (2.01)	2.86 (1.96)
(H/Hmax)	0.52 (0.79)	0.90 (0.92)	0.92 (0.89)

Rr=species richness, H=diversity index, H/Hmax=evenness index.

Despite the differences among the indices with respect to the richness, the tendency with respect to diversity and the

evenness is similar for both techniques. That is why the authors consider that TRFLP is more useful for diversity and dynamic evaluation and Ion Torrent is used for deeper identification until species level.

Conclusions

This study provides a comprehensive morphological and molecular characterization of the efficient microorganisms (EM16) and sugarcane and bamboo soil samples. The diversity, dynamic and taxonomic identification achieved in this study for the EM16 showed the perspective for using these consortia for bioremediation, considering the wide metabolic pathway including the presence of key species like *Lactobacillus* sp. with high potentials for biodegradation of lignocellulosic resilient compounds. Future studies may be aimed at evaluating the mixture of these microorganisms to obtain a microbial consortium with a higher metabolic capacity and increased effectivity in its agricultural or environmental use compared to EM already studied. The use of microbiological and molecular tools under polyphasic approaches allows the completed characterization of non-cultivable microorganisms reported for the first time from an efficient microorganism consortium.

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Conflict of interest statement

The authors declare that there is no conflict of interests regarding the publication of this article.

Author's contributions

JJ and CH formulated the overarching research goals and coordinated the research activity planning and execution. DH and PA provided the study materials, reagents, laboratory samples, instrumentation, computing resources, and other analysis tools. ACP carried out the laboratory activities, applied statistical, mathematical, computational, and other formal techniques to analyze study data. ACP and JJ conducted the research and investigation process, specifically performing the experiments. ZG verified the overall replication/reproducibility of results. All authors

contributed to the writing of the manuscript and carried out the revision of the manuscript.

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